ABUNDANCE

by

Jiaxin Luo

Submitted in partial fulfillment of the requirements for the degree of Master of Science
at
Dalhousie University
Halifax, Nova Scotia
December 2020
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Dedicated to myself, my parents and everyone else who made this possible. Thanks a million!

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#### Abstract

The longline halibut survey is a joint project between the Atlantic Halibut Council (AHC) and Fisheries and Oceans Canada (DFO). This annual survey is used to monitor the status of Atlantic Halibut (Hippoglossus hippoglossus) on the Scotian Shelf and southern Grand Banks (NAFO Divisions 3NOPs4VWX5Zc). The purpose of this thesis is to develop novel statistical analyses of Atlantic Halibut longline survey data for more accurate and precise indices of relative abundance. Reproduction of the survey indices has been completed using hook occupancy data collected between 2017 and 2019 and current hook competition model. In order to reflect spatial patterns in these data, a spatial random field is introduced into the model. Covariates for depth, temperature and area strata (along with their potential interactions) are also introduced as these may affect the survey indices. Model selection has been performed and the selected spatial model is then used to estimate the relative abundance indices for target and non-target species at each survey station for 2017 to 2019. Finally, the estimated indices are aggregated over space using three different approaches to compare the estimated changes in survey indices with those obtained with the nonspatial model currently being used by DFO. The results indicate that the relative abundance indices for halibut are higher on shallow banks in southwest Nova Scotia (NAFO 4X and 4W) and along the shelf edge throughout the management unit.


## Acknowledgements

I would like to express the deepest appreciation to my supervisors Bruce Smith and Joanna Mills Flemming for their persistent support. They guided me through the thesis research, contributed in countless ways to my development and encouraged me all the time.

A special thanks to Brendan Wringe and Nell den Heyer for providing patient and timely responses to all of my questions about the database and the survey. I would also like to thank Zeliang Wang for providing the estimated bottom temperature data.

I would like to thank Yuan (Sofie) Yan for answering all my questions about Spatial Statistics with patience and Ethan Lawler for helping me code with R package TMB.

Last but not the least, I would like to acknowledge the rest of my thesis committee for spending time reading my thesis and providing valuable comments on this thesis.

## Chapter 1

## Introduction

We introduce the annual Atlantic Halibut longline survey and describe the survey data in the first two sections of this chapter. We then discuss the models and R packages that are currently used to obtain the survey indices.

### 1.1 Background



Figure 1.1: Photograph of an Atlantic Halibut [2].

Atlantic Halibut is an economically important demersal flatfish species that is distributed widely throughout the North Atlantic Ocean, as well as being found in parts of the Arctic Ocean. It is commonly found in deep-water channels running between banks and along the edge of the continental shelf. The depth distribution of Atlantic

Halibut ranges from less than 50 m to over 1000 m but they are most abundant in waters from 200 m to 500 m deep [1]. There are two Atlantic Halibut management units in Canadian waters, namely the Scotian Shelf and southern Grand Banks (North Atlantic Fisheries Organization [NAFO] Divisions 3NOPs4VWX5Zc), and the Gulf of St. Lawrence (4RST). This analysis will focus on the 3NOPs4VWX5Zc stock (Figure 1.2).

Together, the AHC and DFO have used an annual longline survey to monitor Atlantic Halibut exploitable biomass since 1998. Longline fishing is a commercial fishing technique that enables massive fish yields and the lenght of the longline for halibut survey is approximately 5 km . Longlines can be placed at the sea surface (pelagic longlines) or at the seafloor (demersal longlines) and consist of a very long mainline. Baited hooks, used to lure and capture target species, are attached to the mainline by branchlines. The survey is completed by commercial fishermen with at-sea on board observers collecting data. The survey was originally stratified into areas of Low, Medium and High catch based on data from commercial fishing logs (1995-1997) [11][29]. Stratified estimates were used until the assessment by Trzcinski et al.[26]. In 2009, a standardized catch rate, calculated from a negative binomial (NB) generalized linear model (GLM), replaced the stratified estimate of mean weight per standard longline set [11][26]. Stratified random design reduces the variance in the mean of the estimate, therefore it was decided to transition to a stratified random survey design with more standardized protocols and expanded geographic coverage in 2016.

The simple stratified mean adjusted catch rate and the NB GLM both assume that halibut are the only species being caught by the longline hooks without accounting for other species competing for hooks. In addition, these methods implicitly assume that all hooks that do not have halibut would still have been able to catch halibut had there been more of these fish in the area. However, some number of hooks will be occupied by species other than halibut and other hooks will be empty with bait still attached or missing. Catch Per Unit Effort (CPUE) ignores the effects of competition
for baited hooks within and between species and the problem of gear saturation [13]. Competition between target and non-target fish for the same hooks may result in gear saturation when the local population density of the target or non-target species exceeds the number of hooks which results in decreased efficiency of fishing effort [5]. It will also lead to bias in the abundance trends estimated using CPUE.

Target and non-target species will respond differently to temperature and depth and this may impact the catchability of longline gear. Over the scale of this large management unit (app 2000km long) temperature and depth will be associated with changes in fish assemblages thereby influencing hook competition. Temperature is also linked to the activity and feeding of ectotherms - which includes fish, amphibians and reptiles whose regulation of body temperature depends on external sources. In summary, there can be spatial and temporal variability in the abundance of target and non-target species and their catchability that could bias an index of abundance that ignores these sources of variability.

Stephen Smith [24] recommended using a multinomial model to account for the number of halibut caught, the number of hooks occupied by other species as well as empty-baited and empty-unbaited hooks and developed a package (llsurv) using the software $\mathrm{R}[21][23]$. The llsurv package uses explicit formulae to compute maximum likelihood estimates of the multinomial model parameters in order to estimate relative abundance.

Here, we reproduce llsurv survey indices using a general purpose optimizer, so that the estimation method will be easier to modify later on. We then propose a novel model framework using Template Model Builder (TMB) [19] for model implementation, and ggplot2 [28] for data visualization from within the R statistical environment [21]. In this new framework, we can include covariates such as depth and temperature. We also include a spatial random field in the model, evaluate the effectiveness of the stratification scheme, and test an alternative and more flexible method for reducing variance in estimates of abundance.

### 1.2 Data Description

The stratified random survey area is divided into $5(4 \mathrm{X} 5 \mathrm{YZ}, 4 \mathrm{~W}, 4 \mathrm{~V}, 3 \mathrm{P}, 3 \mathrm{NO})$ area strata each with 3 (30-130 m, 131-250 m, 251-750 m) depth zones. Depth strata are selected based on exploratory analyses of catch rates by depth using fixed-stations and commercial index sets from the Atlantic Halibut survey. The depth stratum boundaries ( $30-750 \mathrm{~m}$ ) were chosen because they contained most of the fishing sets and inferred Atlantic Halibut habitat. The depth stratification can also provide a proxy for temperature and some bottom habitat information [10]. For example, Smith [24] found that there is a strong correlation between depth and temperature for depths deeper than 280 m for sets in NAFO areas 4VWX. It is worthwhile to mention that the temperature for each longline set is reasonably constant during the fishing but depth can be variable for sets along a slope. In total there are 15 areadepth strata. Stations are randomly assigned to strata with the number of stations allocated proportionally by the size of each stratum. A total of 150 stations were allocated in 2017 and 153 stations were allocated in 2018, while in 2019 additional stations were added to strata with only 2 stations in an effort to reduce the probability of unfished strata or strata with only one station fished.


Figure 1.2: The survey area for the 3NOPs4VWX5Zc Atlantic Halibut stock (with five area strata: $4 \mathrm{X} 5 \mathrm{YZ}, 4 \mathrm{~W}, 4 \mathrm{~V}, 3 \mathrm{P}$ and 3 NO ).

The Atlantic Halibut longline survey with a stratified random design generated the dataset considered here. Trained observers monitored the longline and recorded the conditions of a subset 300 hooks as they were retrieved. Specifically, each longline placed at the seafloor contained 1000 baited hooks and the hook conditions of 10 samples of 30 hooks were recorded for estimating hook occupancy. In addition, there exists a protocol to refish a station if the corresponding set is hauled back with no baited hooks. For each set we obtain counts of empty-unbaited hooks, empty-baited hooks, broken hooks, hooks with target catch and hooks with non-target catch. The corresponding strata information and soak time for each set were also included along with other essential information (Table 1.1 and Figures 1.3 to 1.5). For 2017 and 2018, model estimates of bottom temperature at time and location were obtained from http://marine.copernicus.eu/services-portfolio/access-toproducts/?option=com_csw\&view=details\&product_id=GLOBAL_REANALYSIS_PHY_ 001 _030 [14] by Dr. Zeliang Wang at BIO. The 2019 model estimates were not available for inclusion in this thesis. Issues that arose during the 2019 survey resulted in fewer stations being successfully completed (Table 1.2 and Figure 1.5).

| STATION | STRATUM ID | SOAKMINP3P1 | P1LONG |
| :---: | :---: | :---: | :---: |
| 501 | H11 | 584 | -65.00200 |
| 502 | H11 | 727 | -64.71117 |
| 503 | H11 | 628 | -65.07950 |
| 504 | H11 | 370 | -64.78967 |
| 505 | H11 | 360 | -65.17850 |
| P1LAT | P1DEPTH | EMPTY UNBAITED | EMPTY BAITED |
| 45.45167 | 38 | 229 | 32 |
| 45.22783 | 46 | 245 | 15 |
| 45.14717 | 72 | 159 | 103 |
| 43.77800 | 32 | 37 | 257 |
| 43.34200 | 148 | 93 | 202 |
| MISS BROKEN | HALIBUT | OTHER SPECIES | HOOKS SAMPLED |
| 1 | 0 | 38 | 300 |
| 3 | 0 | 37 | 300 |
| 1 | 6 | 31 | 300 |
| 0 | 1 | 5 | 300 |
| 0 | 5 | 0 | 300 |
| STRATUM ID ALT | MED TEMP | GLORYS TEMP |  |
| 1 | 11.37 | 7.18 |  |
| 1 | 6.72 | 6.72 |  |
| 1 | 10.13 | NaN |  |
| 1 | 3.25 | 4.34 |  |
| 1 | 6.24 | 6.24 |  |

Table 1.1: Partial longline survey data for 2017. STATION indicates the survey station number. STRATUM ID and STRATUM ID ALT both indicate the corresponding stratum id (H11: area strata 4 X 5 YZ and depth strata $30-130 \mathrm{~m}$ ) for the survey station. SOAKMINP3P1, P1LONG and P1LAT represent the soak time and recorded latitude and longitude coordinate for the starting point of each longline set. Columns EMPTY UNBAITED, EMPTY BAITED, MISS BROKEN, HALIBUT and OTHER SPECIES are the count of hooks without bait, with bait, missing or broken, with halibut and other species at retrieval. HOOKS SAMPLED is the number of hooks sampled from the longline set. MED TEMP is the average winter bottom temperature and GLORYS TEMP contains temperature data obtained from Copernicus Marine Environment Monitoring Service (CMEMS) of the European Union (EU). NaN indicates a missing temperature.


Figure 1.3: Locations of the 150 survey stations in 2017. Yellow points are stations 523 and 525 where only 30 hooks were sampled.


Figure 1.4: Locations of the 153 survey stations in 2018.


Figure 1.5: Locations of the 153 survey stations in 2019. Red points are the successfully completed stations and yellow points are the uncompleted stations due to issues that arose during the 2019 survey.

| Stratum ID | Allocation of <br> Survey Stations <br> for 2017 | Allocation of <br> Survey Stations <br> for 2018 | Allocation of <br> Survey Stations <br> for 2019 |
| :---: | :---: | :---: | :---: |
| H11 | $16(16)$ | $16(16)$ | $16(16)$ |
| H12 | $11(11)$ | $11(11)$ | $11(11)$ |
| H13 | $2(2)$ | $3(3)$ | $3(3)$ |
| H21 | $15(15)$ | $15(15)$ | $15(15)$ |
| H22 | $7(7)$ | $7(7)$ | $7(7)$ |
| H23 | $2(2)$ | $3(3)$ | $3(3)$ |
| H31 | $11(11)$ | $11(11)$ | $11(11)$ |
| H32 | $5(5)$ | $5(5)$ | $5(5)$ |
| H33 | $8(8)$ | $8(8)$ | $8(8)$ |
| H41 | $13(13)$ | $13(13)$ | $10(13)$ |
| H42 | $8(8)$ | $8(8)$ | $5(8)$ |
| H43 | $8(8)$ | $8(8)$ | $4(8)$ |
| H51 | $39(39)$ | $39(39)$ | $25(39)$ |
| H52 | $2(2)$ | $3(3)$ | $2(3)$ |
| H53 | $3(3)$ | $3(3)$ | $2(3)$ |
| Total Number of <br> Survey Stations | $150(150)$ | $153(153)$ | $127(153)$ |

Table 1.2: The planned (in brackets) and completed allocation of survey stations per stratum and the total number of survey stations.

For stations 523 and 525 (with stratum id H12) in 2017 (Figure 1.3), the observer only recorded data for 30 hooks which was against observer protocol. Since there are 1000 hooks deployed per line and 10 samples of 30 hooks ( 300 hooks per station) need to be used to estimate hook occupancy, these two stations were data deficient. There are 11 stations in strata H12 (area 4X5YZ with depth 131-250 m) which suggests that removing these two incompletely observed stations will not have a great effect on subsequent analyses (Table 1.2). Therefore, stations 523 and 525 will be dropped from the 2017 analyses.

The missing depths for stations 570, 572 and 573 in 2018 and stations 570, 571, 572, 573 and 574 in 2019 were replaced by the depth estimated using publicly available bathymetric data from the United States National Atmospheric and Oceanographic Administration (NOAA). Depths were extracted from the bathymetric layer using the function getNOAA.bathy from the R package marmap [20] and the coordinate bounds of the survey area were used to import bathymetric data from the NOAA server. Then the depth of the area of interest was obtained from the imported bathymetric data using the get.depth function from the R package marmap. The bathymetric method is used because it is accepted as the best available measurement when depth is not directly measured or recorded during setting of the longline by the onboard observer. $62 \%$ of the absolute differences between survey depth and estimated depth are within 10 meters (Figure 1.6 through Figure 1.8). Stations with large depth variation ( $>|10|$ $m)$ are in areas where there are substantial depth slopes and these rapid depth changes are not well described by the bathymetric models (Figure 1.9 through Figure 1.11).


Figure 1.6: Absolute differences (m) between survey depth and depth obtained using the bathymetric method for 2017.


Figure 1.7: Absolute differences (m) between survey depth and depth obtained using the bathymetric method for 2018 without stations 570, 572 and 573.


Figure 1.8: Absolute differences ( m ) between survey depth and depth obtained using the bathymetric method for 2019 without stations $570,571,572,573$ and 574 .


Figure 1.9: Locations of the survey stations with indication of absolute difference between survey depth and bathymetric depth for 2017. Red points are stations with absolute difference greater than 10 m .


Figure 1.10: Locations of the survey stations with indication of absolute difference between survey depth and bathymetric depth for 2018. Red points are stations with absolute difference greater than 10 m .


Figure 1.11: Locations of the survey stations with indication of absolute difference between survey depth and bathymetric depth for 2019. Red points are stations with absolute difference greater than 10 m .

From 2017 to 2019, $70 \%$ to $90 \%$ of the sets in the longline survey have temperature data derived from attached VEMCO minilogs (Table 1.3). The model generated bottom temperature data (GLORYS12V1) at time and location were used for survey stations which did not have temperature data from the minilogs. The GLORYS12V1 from Copernicus Marine Environment Monitoring Service (CMEMS) of the European Union (EU) is an eddy-resolving global ocean product, which assimilates observational data (e.g. satellite sea surface temperature data, satellite altimeter data, temperature and salinity data from ARGO profilers and ship surveys). The GLORYS12V1 reanalysis product has a horizontal resolution of $1 / 12^{\circ}$, and 50 vertical levels [14]. The model estimated bottom temperature is used because the observed temperature and generated temperature are well correlated (Figure 1.12). Though the model tends to overestimate or underestimate the temperatures near Shelf edges, the residuals between the observed and estimated temperatures are small (Figures 1.13 to 1.15).

| Year | Number of Stations <br> with Temperature | Number of Stations <br> without Temperature | Percentage of Stations <br> with Temperature |
| :---: | :---: | :---: | :---: |
| 2017 | 134 | 14 | $90.54 \%$ |
| 2018 | 125 | 28 | $81.70 \%$ |
| 2019 | 101 | 16 | $79.53 \%$ |

Table 1.3: Summary of the number of stations with and without observed temperature data and the percentage of stations with observed temperature.


Figure 1.12: Observed temperature and model estimated temperature for 2017 and 2018.


Figure 1.13: The residuals between observed temperature and model estimated temperature for 2017 and 2018.


Figure 1.14: The residuals between observed temperature and model estimated temperature for 2017. The small black points are the stations with missing temperature.


Figure 1.15: The residuals between observed temperature and model estimated temperature for 2018. The small black points are the stations with missing temperature.

It can be noticed from Table 1.1 that soak times in the survey have varied over survey stations. However from Figure 1.16 we can observe that the distribution of soak time does not vary over years. Comparing to the other depth strata, there are more longline sets with soak time below 450 min in depth strata $30-130 \mathrm{~m}$ (Figure 1.17). The distribution of soak time in different area strata varies but there is no obvious pattern for soak time in different area strata (Figure 1.18).


Histogram of Soak Time for 2018


Histogram of Soak Time for 2019


Figure 1.16: Histograms of the soak time by year for 2017 to 2019.














Figure 1.18: Histograms of the soak time by area for 2017 to 2019.




### 1.3 Survey Index Reproduction

### 1.3.1 Multinomial Exponential Model

The multinomial exponential model (MEM) originally proposed by Rothschild [22] and modified by Etienne et al. [13] is currently used by DFO for the Atlantic Halibut longline survey.

Let $T_{T}$ be the time of catching a target species (halibut) on a hook and assume $T_{T}$ follows an exponential distribution with rate $\lambda_{T}$. Then the probability of catching a halibut before time $u$ is

$$
P\left(T_{T} \leq u\right)=1-e^{-\lambda_{T} u}
$$

Let $T_{N T}$ be the time it takes for a non-target species to be caught and assume $T_{N T}$ follows an exponential distribution with rate $\lambda_{N T}$. The probability distribution of the time it takes for either a target or non-target species to be caught, $T=\min \left\{T_{T}, T_{N T}\right\}$ is also exponential with rate $\lambda=\lambda_{T}+\lambda_{N T}$.

The possible states for a hook after soak time $S$ upon retrieval onboard are,

- $\{I=0\}=\{$ The hook is still baited $\}$ with probability

$$
P(I=0)=P(T>S)=e^{-\lambda S}
$$

- $\{I=$ Target $\}=\{$ The hook has a target species $\}$ with probability

$$
P(I=\text { Target })=P\left(T_{T}<T_{N T} \mid T<S\right) P(T<S)=\frac{\lambda_{T}}{\lambda}\left(1-e^{-\lambda S}\right)
$$

- $\{I=N$ Target $\}=\{$ The hook has a non-target species $\}$ with probability

$$
P(I=N \text { Target })=P\left(T_{N T}<T_{T} \mid T<S\right) P(T<S)=\frac{\lambda_{N T}}{\lambda}\left(1-e^{-\lambda S}\right)
$$

where I is an indicator variable.
Assuming that all the hooks on a longline are independent, the likelihood for hooks being baited or not baited and if not baited containing a halibut or another
species is given by

$$
\begin{aligned}
L\left(\lambda_{T}, \lambda_{N T}\right) & =\binom{N}{N_{B}}\binom{N_{T}+N_{N T}}{N_{T}}\left(e^{-\lambda S}\right)^{N_{B}} \\
& \times\left(1-e^{-\lambda S}\right)^{N_{T}+N_{N T}}\left(\frac{\lambda_{T}}{\lambda}\right)^{N_{T}}\left(\frac{\lambda_{N T}}{\lambda}\right)^{N_{N T}}
\end{aligned}
$$

where

- N is the number of hooks on the longline,
- $N_{B}$ is the number of baited hooks at the end of the soak time,
- $N_{T}$ is the number of the target species caught,
- $N_{N T}$ is the number of the non-target species caught.

This model is called the Multinomial Exponential Model (MEM) [22]. The expected numbers of the halibut caught by a longline with N hooks is N times the probability $P(I=$ Target $)$ for a halibut being caught when there is competition from other species. Therefore the number of the halibut that were observed reflects an index of the abundance of that species modified by the relative abundance of other species competing for the same hooks. Parameters $\lambda_{T}$ and $\lambda_{N T}$ reflect the relative abundance of target and non-target species respectively.

It is common for some hooks to return with no bait and no fish at the end of soak time in longline fishing. The MEM was modified by Etienne et al. [13] to account for these empty-unbaited hooks so that more precise estimates of annual indices of relative abundance could be obtained. It's assumed that all empty-unbaited hooks are caused by the escape of fish. That is,

- The number $N_{B}$ of baited hooks out of N total hooks at the end of the soak time is binoamially distributed with probability of success $e^{-\lambda S}$ :

$$
N_{B} \sim \mathcal{B}\left(N, e^{-\lambda S}\right)
$$

- Given the number of empty-unbaited hooks $\left(N-N_{B}\right)$, the number of halibut caught is $N_{T}+N_{E}^{(T)}$ and also has a binomial distribution:

$$
N_{T}+N_{E}^{(T)} \left\lvert\, N-N_{B} \sim \mathcal{B}\left(N-N_{B}, \frac{\lambda_{1}}{\lambda}\right)\right.
$$

where $N_{E}^{(T)}$ is the number of caught halibut escaping.

- Given $N_{T}+N_{E}^{(T)}$, the total number of halibut on the longline is $N_{T}$ and is binomially distributed:

$$
N_{T} \mid N_{T}+N_{E}^{(T)} \sim \mathcal{B}\left(N_{T}+N_{E}^{(T)},\left(1-p_{T}\right)\right)
$$

where $p_{T}$ is the escapement rate of caught target fish.

- Given $N_{N T}+N_{E}^{(N T)}$, the total number of non-target species on the longline is $N_{N T}$ and also follows a binomial distribution:

$$
N_{N T} \mid N_{N T}+N_{E}^{(N T)} \sim \mathcal{B}\left(N_{N T}+N_{E}^{(N T)},\left(1-p_{N T}\right)\right)
$$

where $N_{E}^{(N T)}$ is the number of caught non-target species escaping and $p_{N T}$ is the escaping rate of caught non-target fish.
$N_{E}^{(N T)}$ and $N_{E}^{(T)}$ are missing quantities but the sum $N_{E}$ of these two quantities is observed. Therefore the likelihood function of the model is:

$$
\begin{aligned}
L\left(\lambda_{T}, \lambda_{N T}, p_{T}, p_{N T}\right) & =\frac{N!}{N_{B}!N_{T}!N_{N T}!N_{E}!}\left(e^{-\lambda S}\right)^{N_{B}}\left(1-e^{-\lambda S}\right)^{N-N_{B}} \\
& \times\left(\frac{\lambda_{T}}{\lambda}\left(1-p_{T}\right)\right)^{N_{T}}\left(\frac{\lambda_{N T}}{\lambda}\left(1-p_{N T}\right)\right)^{N_{N T}}\left(\frac{\lambda_{T} p_{T}+\lambda_{N T} p_{N T}}{\lambda}\right)^{N_{E}} .
\end{aligned}
$$

This is a multinomial distribution with parameters $N$ and $\alpha$ :

$$
\left(N_{B}, N_{T}, N_{N T}, N_{E}\right) \sim \mathcal{M}(N, \alpha)
$$

where N is the number of hooks on a longline set and $\alpha=\left(e^{-\lambda S},\left(1-e^{-\lambda S}\right) \frac{\lambda_{T}}{\lambda}(1-\right.$ $\left.\left.p_{T}\right),\left(1-e^{-\lambda S}\right) \frac{\lambda_{N T}}{\lambda}\left(1-p_{N T}\right),\left(1-e^{-\lambda S}\right) \frac{\lambda_{T} p_{T}+\lambda_{N T} p_{N T}}{\lambda}\right)$.

There were two models proposed by Etienne et al. [13] based on two different situations. The first model (MEM1) assumed that empty-unbaited hooks were caused by non-target species taking the bait and escaping with the corollary that the probability of escape for target species would be $0\left(p_{T}=0\right)$. The second model (MEM2) assumed that empty-unbaited hooks came from either target or non-target species. In other words, the escaping rate is the same for target and non-target catch $\left(p_{T}=p_{N T}\right)$. However, those authors who have investigated the impact of bait loss on longline
estimates have generally assumed that missing baits were caused by non-target fish [7][17][27]. Since the longline survey is designed for catching and retaining the target species in terms of bait, hook type and size, depth, and timing, non-capture would be more likely for other species in the area. In addition, the target species would also have lower abundance than all other possibly caught species. Moreover, assigning the empty-unbaited hooks to species other than halibut will at worst lead to an underestimation of the halibut count [13]. Therefore, MEM1 was selected and the likelihood function was re-written as follows:

$$
\begin{aligned}
L\left(\lambda_{T}, \lambda_{N T}, p_{N T}\right) & =\frac{N!}{N_{B}!N_{T}!N_{N T}!N_{E}!}\left(e^{-\lambda S}\right)^{N_{B}}\left(1-e^{-\lambda S}\right)^{N-N_{B}} \\
& \times\left(\frac{\lambda_{T}}{\lambda}\right)^{N_{T}}\left(\frac{\lambda_{N T}}{\lambda}\left(1-p_{N T}\right)\right)^{N_{N T}}\left(\frac{\lambda_{N T} p_{N T}}{\lambda}\right)^{N_{E}}
\end{aligned}
$$

The maximum likelihood estimators of MEM1 with a constant soak time are given by:

$$
\begin{gather*}
\hat{\lambda}_{T}=\frac{N_{T+}}{N_{+}-N_{B+}} \frac{1}{S} \log \left(\frac{N_{+}}{N_{B+}}\right)  \tag{1.1}\\
\hat{\lambda}_{N T}=\frac{N_{N T+}+N_{E+}}{N_{+}-N_{B+}} \frac{1}{S} \log \left(\frac{N_{+}}{N_{B+}}\right)  \tag{1.2}\\
\hat{p}_{N T}=\frac{N_{E+}}{N_{E+}+N_{N T+}} \tag{1.3}
\end{gather*}
$$

where $N_{B+}=\sum_{l=1}^{L} N_{B_{l}}, N_{T+}=\sum_{l=1}^{L} N_{T_{l}}, N_{N T+}=\sum_{l=1}^{L} N_{N T_{l}}, N_{E+}=\sum_{l=1}^{L} N_{E_{l}}$ and $l$ is the number of longline sets. Additionally, empty-unbaited hooks also include broken and missing hooks.

### 1.3.2 Template Model Builder (TMB)

We coded the MEM1 proposed by Etienne et al. [13] in TMB by first constructing the negative log-likelihood function in R. After ensuring that the negative log-likelihood function in R was working appropriately, the function was converted into $\mathrm{C}++$ code and the TMB package was used for estimation so that the MEM1 could be efficiently implemented using simple $\mathrm{C}++$ templates. The model was formulated in $\mathrm{C}++$ and
the data were manipulated in R . The function MakeADFun from the R package TMB was used to construct objective functions and pass the objective function value to the optimizer in R for the final estimation. This package is used because it combines external libraries CppAD, Eigen and CHOLMOD to obtain an efficient implementation of the applied Laplace approximation with exact derivatives and can speed up a process significantly when the number of iterations is high. For more details about the TMB package please refer to the TMB reference manual by Kristensen et al. [18]. In addition, R codes used for analysis and data visualization are provided in the Appendices.

### 1.3.3 Comparison of Results

The R package llsurv [24] uses explicit formulae for maximum likelihood estimates. These only allow one value for soak time therefore the mean of the soak time for the survey was used to reproduce results. The identical results obtained using two different methods and packages (Table 1.4) indicate that survey indices estimated using the explicit method can be successfully reproduced using a general purpose optimizer. For the year 2017, the estimated relative abundance indices with soak time fixed at 500.5068 minutes for the target $\left(\hat{\lambda}_{T}\right)$ and non-target species $\left(\hat{\lambda}_{N T}\right)$ are $1.843592 \mathrm{e}-05$ and 0.001904881 respectively. The estimated survey indices with constant soak time ( 478.1895 minutes) for the target $\left(\hat{\lambda}_{T}\right)$ and non-target species $\left(\hat{\lambda}_{N T}\right)$ are $2.661856 \mathrm{e}-05$ and 0.002068632 respectively in 2018 . The probability of escape for the non-target species $\left(\hat{p}_{N T}\right)$ is 0.8908173 . For 2019, the estimated relative abundance indices with fixed soak time ( 463.2283 minutes) for the target $\left(\hat{\lambda}_{T}\right)$ and non-target species $\left(\hat{\lambda}_{N T}\right)$ are $2.324046 \mathrm{e}-05$ and 0.002227046 respectively.

The TMB generated indices of abundance can also be estimated with set-specific soak time (Table 1.5). The survey indices estimated using the set-specific soak time will be more accurate as the Multinomial model assumes that catch increases linearly with soak time. However error rates between the results obtained using fixed soak time and set-specific time are small. The error rate is obtained using the difference
between the survey indices estimated with fixed soak time and set-specific soak time and divided by the survey index estimated with set-specific soak time. Specifically, error rates between the estimated survey indices for target species with mean of soak time and set-specific soak time are approximately $1.90 \%$ in $2017,0.37 \%$ in 2018 and $0.23 \%$ in 2019. A comparison of the proportion of hooks with halibut from the survey with soak time indicates that soak time does not appear to have any effect on the proportion of halibut caught. A similar conclusion can be drawn from the proportion of hooks with catches of the other non-target species as well (Figure 1.19).

| Year | Mean of the <br> Soak Time (mins) | $\hat{\lambda}_{T}$ (llsurv) | $\hat{\lambda}_{N T}$ (llsurv) | $\hat{\lambda}_{T}$ (TMB) | $\hat{\lambda}_{N T}$ (TMB) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 500.5068 | $1.843592 \mathrm{e}-05$ | 0.001904881 | $1.843592 \mathrm{e}-05$ | 0.001904881 |
| 2018 | 478.1895 | $2.661856 \mathrm{e}-05$ | 0.002068632 | $2.661856 \mathrm{e}-05$ | 0.002068632 |
| 2019 | 463.2283 | $2.324046 \mathrm{e}-05$ | 0.002227046 | $2.324046 \mathrm{e}-05$ | 0.002227046 |

Table 1.4: The estimated relative abundance indices across the entire survey area with mean of the soak time for the survey for target and non-target species using $R$ packages TMB and llsurv for hook occupancy data from the halibut longline survey for the years 2017, 2018 and 2019.

| Year | $\hat{\lambda}_{T}$ | $\hat{\lambda}_{N T}$ | $\hat{p}_{N T}$ |
| :---: | :---: | :---: | :---: |
| 2017 | $1.878635 \mathrm{e}-05$ | 0.001941087 | 0.8795759 |
| 2018 | $2.671808 \mathrm{e}-05$ | 0.002076371 | 0.8908173 |
| 2019 | $2.329181 \mathrm{e}-05$ | 0.002232086 | 0.8963697 |

Table 1.5: The estimated relative abundance indices across the entire survey area with set-specific soak time for target and non-target species and the escaping probability of caught non-target species using the TMB package for hook occupancy data from the halibut longline survey for the years 2017, 2018 and 2019.

Figure 1.19: A comparison of the proportion of halibut and non-target species caught with different soak times for 2017, 2018
and 2019.

## Chapter 2

## Spatial Models and Variable Selection

Here we start with a discussion of the spatial patterns found in the survey data. The following two sections then focus on the application of spatial statistics to the current survey indices estimation model and culminate with the proposal of a novel model formulation. Variable selection for this model is conducted in the final section.

### 2.1 Spatial Patterns

Typically when a region exhibits spatial structure, the points closer to each other are more strongly correlated than the distant points. The maps of the proportion of baited and unbaited hooks in each set show obvious spatial patterns (Figures 2.1 to 2.3). Here unbaited hooks include missing and broken hooks as well as hooks that came back with halibut, other species or without bait. There are more unbaited hooks in the depth zone $251-750 \mathrm{~m}$ and more baited hooks in the depth zone 30-130 m . With regard to the area strata, there are more unbaited hooks in areas 4X5XY, 4 W and 3 P . This result is further supported by ANOVA tests of the proportion of unbaited hooks by strata (Tables 2.1 to 2.3 ). The p-values ( $9.4 \mathrm{e}-08$ and $9.81 \mathrm{e}-07$ ) for these tests are very small indicating that there are significant differences in the proportion of unbaited hooks by strata. To attempt to deal with this spatial structure a spatial random field is considered. A spatial random field is a random function over an arbitrary domain meaning that it is a function $f(x)$ that takes on a random value at each point $x \in \mathbb{R}^{n}$. It is also defined as a generalization of a stochastic process taking values in a Euclidean space, and defined over a parameter space of at least one dimension [3]. Abundance estimates obtained with inclusion of a spatial random field will be compared to those previously obtained based on the stratification scheme.


Figure 2.1: The distribution of unbaited hooks in 2017. Here unbaited hooks include missing and broken hooks as well as hooks that came back with halibut, other species or without bait.


Figure 2.2: The distribution of unbaited hooks in 2018. See caption for Figure 2.1.


Figure 2.3: The distribution of unbaited hooks in 2019. See caption for Figure 2.1.

|  | DF | sum Sq | Mean Sq | F value | $\operatorname{Pr}(>\mathrm{F})$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| STRATUM ID | 14 | 4.452 | 0.3180 | 4.323 | $2.77 \mathrm{e}-06$ |  |
| Residuals | 133 | 9.783 | 0.0736 |  |  |  |

Table 2.1: Summary of the ANOVA test of proportion of unbaited hooks by strata for 2017.

|  | DF | sum Sq | Mean Sq | F value | $\operatorname{Pr}(>\mathrm{F})$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| STRATUM ID | 14 | 4.850 | 0.3465 | 6.693 | $2.98 \mathrm{e}-10$ |  |
| Residuals | 138 | 7.143 | 0.0518 |  |  |  |

Table 2.2: Summary of the ANOVA test of proportion of unbaited hooks by strata for 2018.

|  | DF | sum Sq | Mean Sq | F value | $\operatorname{Pr}(>\mathrm{F})$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| STRATUM ID | 14 | 2.986 | 0.21328 | 2.92 | 0.000834 |  |
| Residuals | 112 | 8.180 | 0.07304 |  |  |  |

Table 2.3: Summary of the ANOVA test of proportion of unbaited hooks by strata for 2019.

### 2.2 Models with a Gaussian Random Field

### 2.2.1 Binomial Model with Gaussian Random Field (glmmTMB)

The number of unbaited hooks ( $N_{U N B, i}$ ), which includes missing and broken hooks as well as hooks that came back with halibut, other species and without bait, on a longline set at the end of the soak time is binomially distributed

$$
N_{U N B, i} \sim \mathcal{B}\left(N_{i}, \pi_{i}\right) .
$$

The logit of the probability $\pi_{i}$ of obtaining unbaited hooks on a longline set can be modelled as a linear function of soak time $\left(x_{i_{-} s o a k}\right)$, depth $\left(x_{i_{-} \text {depth }}\right)$ and temperature $\left(x_{i_{-t e m p}}\right)$ at each survey location. That is

$$
\begin{equation*}
\operatorname{logit}\left(\pi_{i}\right)=\beta_{0}+\beta_{1} \times x_{\text {soak_i }}+\beta_{2} \times x_{\text {depth_}^{2} i}+\beta_{3} \times x_{\text {temp }-i}, \tag{2.1}
\end{equation*}
$$

and the binomial model with Gaussian random field is defined as

$$
\begin{equation*}
\operatorname{logit}\left(\pi_{i}\right)=\beta_{0}+\beta_{1} \times x_{\text {soak_i }}+\beta_{2} \times x_{\text {depth_i }}+\beta_{3} \times x_{\text {temp_ } i}+\omega_{i} \tag{2.2}
\end{equation*}
$$

where $\omega_{i}$ is a Gaussian random field with Matérn covariance function $C(h)$. A Matérn covariance is generally used to define the statistical covariance between measurements made at two locations that are $h$ units distant from each other. The covariance is stationary as it only depends on distances between locations. It is also isotropic when distance is Euclidean distance. A Matérn covariance function is used because it is flexible and includes both the exponential and Gaussian covariance functions as special cases, moving between them via a smoothness parameter. The Matérn covariance function [16] is defined as

$$
C(h)=\frac{\sigma^{2}}{\Gamma(\nu) 2^{\nu-1}}\left(\frac{h}{\phi}\right)^{\nu} K_{\nu}\left(\frac{h}{\phi}\right),
$$

where

- $K_{\nu}$ : modified Bessel function of the second kind,
- $\sigma^{2}$ : variance/sill parameter,
- $\phi>0$ : range parameter $\Rightarrow$ effective/practical range (the distance at which the correlation decreases to 0.05),
- $\nu>0$ : smoothness parameter $\Rightarrow$ mean square differentiability of the process (Hausdorff/fractal dimension of Gaussian sample paths).
- $h$ : standardized distance matrix of the longline survey location,

Special cases:
$\nu=0.5 \Rightarrow$ Exponential covariance function: $C(h)=\sigma^{2} \exp \left(-\frac{h}{\phi}\right)$
$\nu=\infty \Rightarrow$ Gaussian covariance function: $C(h)=\sigma^{2} \exp \left(-\frac{h^{2}}{\phi^{2}}\right)$
The estimation of $\pi_{i}$ can be performed using package Generalized Linear Mixed Models using Template Model Builder (glmmTMB) [9] as it fits linear and generalized linear mixed models with TMB and has built-in spatial covariance options. glmmTMB fits models using maximum likelihood estimation via TMB and random effects are integrated out using the Laplace approximation.

The probability of getting unbaited hooks was predicted at 1000 randomly selected locations from the block id which is the unique identifier for each grid cell of the entire survey area. The grid is made up of $2 \mathrm{~km} \times 2 \mathrm{~km}$ blocks, with the point indicated being the center of the block. The same 1000 points were used for all three predictions. The predictions were obtained using mean of the soak time for the survey and without taking temperature into account as the soak time and temperature data for the block id are not easily available.

Similar to spatial patterns displayed by observed data (Figures 2.1 through 2.3), the probability of getting unbaited hooks is higher in areas 4X5YZ, 4W and 3P (Figures 2.4 through 2.6). Though glmmTMB has bulit-in spatial covariance options, the family object in function glmmTMB does not contain the multinomial model which allows for estimating the relative abundance indices for halibut directly. The binomial model can only estimate the probability of getting unbaited hooks but the goal is to estimate the survey indices for halibut. Therefore instead of using glmmTMB, the multinomial exponential model (MEM1) with a Gaussian random field will be coded in $\mathrm{C}++$ and used for analysis.


Figure 2.4: The predicted probabilities of getting unbaited hooks at 1000 randomly selected locations for 2017.


Figure 2.5: The predicted probabilities of getting unbaited hooks at 1000 randomly selected locations for 2018.


Figure 2.6: The predicted probabilities of getting unbaited hooks at 1000 randomly selected locations for 2019.

### 2.2.2 Multinomial Exponential Model (MEM1) with Gaussian Random Field (TMB)

The relative abundance index $\lambda_{i}$ at each survey station is affected by the Gaussian random field $\omega_{i}$ with Matérn covariance function, therefore we will modify the MEM1 described in the previous section by adding a random field to both target and nontarget species. The smoothness parameter $\nu$ of the Matérn covariance function is typically fixed at 1 since it is difficult to estimate. The initial value of the range parameter $\phi$ is the area of the standardized coordinates of the survey stations. The relative abundance indices for target and non-target species at each survey station are defined as

$$
\begin{gather*}
\hat{\lambda}_{T_{-i}}=\exp \left(\beta_{T_{\_} 0}+\omega_{T_{-} i}\right)  \tag{2.3}\\
\hat{\lambda}_{N T_{\_} i}=\exp \left(\beta_{N T_{-} 0}+\omega_{N T_{\_} i}\right) \tag{2.4}
\end{gather*}
$$

There are two scenarios for $\omega_{T_{-i}}$ and $\omega_{N T_{-} i}$, the first one is that they are from the same random field (Figure 2.7). This means that $\omega_{T_{-} i}$ and $\omega_{N T_{\_} i}$ share one covariance function, so one choice of parameters $\phi, \sigma$ and $\nu$ evaluated for two different sets of $\omega$. The other one is that $\omega_{T_{-i} i}$ and $\omega_{N T_{-} i}$ are from different random fields meaning that
$\omega_{T_{-} i}$ and $\omega_{N T_{-} i}$ each has a covariance function, so two sets of parameters $\phi, \sigma$ and $\nu$ evaluated for two different sets of $\omega$ (Figures 2.8 and 2.9). By comparing the plots of correlation matrices for target and non-target species under the second situation for 2017 (Figures 2.8 and 2.9), it can be observed that patterns of correlation matrices for target and non-target species are quite different. Therefore, the model with $\omega_{T_{-} i}$ and $\omega_{N T_{-} i}$ from different random fields is used.


Figure 2.7: The correlation matrix for both target and non-target species for 2017. $\omega_{T \_i}$ and $\omega_{N T_{-} i}$ are from same random field.


Figure 2.8: The correlation matrix for target species for 2017. $\omega_{T \_i}$ and $\omega_{N T \_i}$ are from different random fields.


Figure 2.9: The correlation matrix for non-target species for 2017. $\omega_{T_{-} i}$ and $\omega_{N T_{\_} i}$ are from different random fields.

### 2.2.3 Variable Selection

Covariates for depth, temperature and area strata (along with their potential interactions) are also introduced into the model (Equations 2.5 and 2.6). Since these may affect the relative abundance indices of target and non-target species

$$
\begin{gather*}
\hat{\lambda}_{T_{-} i}=\exp \left(\beta_{T_{-} 0}+\omega_{T_{-} i}+\beta_{T_{-}} \times x_{1_{-} i}+\beta_{T_{-}} \times x_{2_{-} i}+\beta_{T_{-} 3} \times x_{1_{-i}} \times x_{2_{-} i}+\ldots\right)  \tag{2.5}\\
\hat{\lambda}_{N T_{-} i}=\exp \left(\beta_{N T_{\_} 0}+\omega_{N T_{-} i}+\beta_{N T_{-} 1} \times x_{1_{-} i}+\beta_{N T_{-} 2} \times x_{2_{-} i}+\beta_{N T_{-} 3} \times x_{1_{-} i} \times x_{2_{-} i}+\ldots\right) . \tag{2.6}
\end{gather*}
$$

From the simplest model (the model without covariates) to the most complex model (the model with temperature, area strata, depth strata and interactions between area and depth strata), there are 14 models to be compared (Tables 2.4 through 2.6). Both Bayesian information criterion (BIC) and Akaike information criterion (AIC) are used and the model is selected based on BIC scores where

$$
B I C=k \ln (n)-2 \ln (L(\hat{\theta}))
$$

and

$$
A I C=2 k-2 \ln (L(\hat{\theta}))
$$

where

- n is the sample size,
- k is the number of parameters in the model plus the intercept,
- $\theta$ is the set of all parameters,
- $L(\hat{\theta})$ represents the maximized likelihood of the model under consideration. AIC and BIC are criteria for model selection, the optimal model is selected based on the minimum AIC or BIC. BIC is used as the criteria for model selection because AIC penalizes the number of parameters less strongly than the BIC meaning that AIC tends to choose model with more parameters which may cause overfitting.

| Model | Minimized Negative <br> Log-likelihood | K | AIC | BIC |
| :---: | :---: | :---: | :---: | :---: |
| With Temperature | 3692.523 | 4 | 7393.046 | 7405.035 |
| With Depth + Temp | 3692.381 | 6 | 7396.762 | 7414.745 |
| With Temp + <br> Depth Strata (3) | 3692.088 | 8 | 7400.177 | 7424.155 |
| Without Covariates | 3712.183 | 2 | 7428.366 | 7434.361 |
| With Area Strata (5) + <br> Depth + Temp | 3683.438 | 14 | 7394.875 | 7436.836 |
| With Depth | 3709.150 | 4 | 7426.300 | 7438.289 |
| With Depth Strata (3) | 3710.374 | 4 | 7428.748 | 7440.737 |
| With Area Strata (5) <br> Depth | 3692.886 | 12 | 7409.771 | 7445.738 |
| With Area Strata (5) <br> Depth Strata (3) + Temp | 3683.370 | 16 | 7398.740 | 7446.696 |
| With Area Strata (5) <br> Depth Strata (3) | 3693.860 | 14 | 7415.719 | 7457.680 |
| With Area Strata (5) * <br> Depth Strata (3) + Temp | 3667.764 | 32 | 7399.527 | 7495.438 |
| With Area Strata (5) * <br> Depth | 3704.438 | 20 | 7448.877 | 7508.821 |
| With Area Strata (5) * <br> Depth Strata (3) | 3679.827 | 30 | 7419.653 | 7509.570 |
| With Area Strata (5) * <br> Depth + Temp | 3786.990 | 22 | 7617.980 | 7683.919 |

Table 2.4: AIC and BIC for different models for 2017. The models are ranked according to their BIC scores in increasing order. Parameter for random field is not included when calculating AIC and BIC. K contains the parameters for both target and non-target species. There are 5 area strata and 3 depth strata. The symbol "*" represents the interactions between the covariates.

| Model | Minimized Negative <br> Log-likelihood | K | AIC | BIC |
| :---: | :---: | :---: | :---: | :---: |
| With Temperature | 3597.365 | 4 | 7202.729 | 7214.851 |
| With Depth + Temp | 3597.191 | 6 | 7206.382 | 7224.565 |
| With Temp + <br> Depth Strata (3) | 3595.191 | 8 | 7206.383 | 7230.626 |
| With Area Strata (5) + <br> Depth + Temp | 3585.981 | 14 | 7199.961 | 7242.387 |
| With Area Strata (5) + <br> Depth Strata (3) Temp | 3584.606 | 16 | 7201.211 | 7249.698 |
| With Depth Strata (3) | 3616.249 | 4 | 7240.497 | 7252.619 |
| Without Covariates | 3621.410 | 2 | 7246.820 | 7252.881 |
| With Depth | 3619.843 | 4 | 7247.685 | 7259.807 |
| With Area Strata (5) + <br> Depth | 3604.551 | 12 | 7233.103 | 7269.468 |
| With Area Strata (5) <br> Depth Strata (3) | 3600.357 | 14 | 7228.714 | 7271.140 |
| With Area Strata (5) * <br> Depth Strata (3) + Temp | 3575.792 | 32 | 7215.584 | 7312.558 |
| With Area Strata (5) * <br> Depth Strata (3) | 3594.379 | 30 | 7248.758 | 7339.671 |
| With Area Strata (5) * <br> Depth | 3637.866 | 20 | 7315.733 | 7376.341 |
| With Area Strata (5) * <br> Depth + Temp | 3656.634 | 22 | 7357.267 | 7423.937 |

Table 2.5: AIC and BIC for different models for 2018. See caption for Table 2.4.

| Model | Minimized Negative <br> Log-likelihood | K | AIC | BIC |
| :---: | :---: | :---: | :---: | :---: |
| Without Covariates | 3447.512 | 2 | 6899.023 | 6904.712 |
| With Depth Strata (3) | 3447.203 | 4 | 6900.565 | 6911.942 |
| With Depth | 3446.283 | 4 | 6902.407 | 6913.784 |
| With Area Strata (5) <br> Depth | 3437.861 | 12 | 6899.723 | 6933.853 |
| With Area Strata (5) <br> Depth Strata (3) | 3436.680 | 14 | 6901.360 | 6941.178 |
| With Area Strata (5) * <br> Depth | 3439.862 | 20 | 6919.724 | 6976.607 |
| With Area Strata (5) * <br> Depth Strata (3) | 3422.570 | 30 | 6905.139 | 6990.465 |

Table 2.6: AIC and BIC for different models for 2019. See caption for Table 2.4.

We can observe from Tables 2.4 through 2.6 that the model to be selected is the model with only temperature which is defined as

$$
\begin{gather*}
\hat{\lambda}_{T_{-} i}=\exp \left(\beta_{T \_0}+\beta_{T_{-} 1} \times x_{t e m p \_i}+\omega_{T \_i}\right)  \tag{2.7}\\
\hat{\lambda}_{N T_{-} i}=\exp \left(\beta_{N T \_0}+\beta_{N T-1} \times x_{t e m p-i}+\omega_{N T \_i}\right) . \tag{2.8}
\end{gather*}
$$

However, temperature will not be included in the model used to provide an index of abundance because it is not possible to estimate the temperature data for sets that did not have temperature collected in 2019 at this time (Table 1.3). Additionally, the differences between the survey indices for halibut estimated by models with and without temperature are small. This indicates that while temperature is statistically significant it does not have a great impact on the predictions (Figures 2.10 and 2.11). By comparing the BIC scores of the models in Tables 2.4 to 2.6 the model with the lowest BIC and without temperature is the model without covariate. The model is then reduced to

$$
\begin{gather*}
\hat{\lambda}_{T_{-} i}=\exp \left(\beta_{T_{-} 0}+\omega_{T_{-} i}\right)  \tag{2.9}\\
\hat{\lambda}_{N T_{-} i}=\exp \left(\beta_{N T_{-} 0}+\omega_{N T_{-} i}\right) . \tag{2.10}
\end{gather*}
$$

It also can be observed from Tables 2.4 to 2.6 that strata are not statistically important in the spatial model.


Figure 2.10: Relative abundance indices for halibut and corresponding standard errors estimated by models with and without temperature for 2017.


Figure 2.11: Relative abundance indices for halibut and corresponding standard errors estimated by models with and without temperature for 2018.

## Chapter 3

## Estimated Survey Indices

Here we discuss the relative abundance indices for halibut estimated using our spatial model and how these may be aggregated to obtain annual survey indices.

### 3.1 Estimated Survey Indices with Corresponding Standard Errors

It can be observed from Figures 3.1 and 3.2 that the estimated survey indices for halibut are higher in the shallow areas. There are also more halibut in the survey area 4 X 5 YZ with depth ranges from 30 to 130 m for 2017, 2018, and 2019. In comparison to 2017, both 2018 and 2019 have more halibut in survey area 4 W with depth ranges from 30 to 130 m . Additionally, compared to 2018, both 2017 and 2019 have less halibut in survey area 3P. The highest estimated relative abundance indices of the target species appear in survey areas 4V, 4X5YZ and 4W for 2017, 2018 and 2019 respectively. From Figures 3.3 to 3.5, we can observe that the error bars for larger estimated survey indices are wider.


Figure 3.1: Estimated relative abundance indices for halibut for 2017 (top), 2018 (middle) and 2019 (bottom) based on the multinomial exponential model with Gaussian random field.



Figure 3.3: Estimated relative abundance indices and corresponding standard errors for halibut for 2017 based on the multinomial exponential model with Gaussian random field.


Figure 3.4: Estimated relative abundance indices and corresponding standard errors for halibut for 2018 based on the multinomial exponential model with Gaussian random field.


Figure 3.5: Estimated relative abundance indices and corresponding standard errors for halibut for 2019 based on the multinomial exponential model with Gaussian random field.

### 3.2 Interpolation of Estimated Relative Abundance Indices and Random Field

The spatial model can be used to estimate the survey indices for target and non-target species at the specific survey stations but we would also like to predict the relative abundance indices over the entire survey area, therefore kriging is used. Kriging [16] is a method of interpolation by which the interpolated values are modeled by a Gaussian process governed by prior covariances. The relative abundance indices are interpolated over the block id as the survey locations are selected from the block id. The Gaussian random field $\omega$ is interpolated using the formula

$$
\hat{Z}\left(\mathbf{s}^{*}\right)=E\left\{Z\left(\mathbf{s}^{*}\right) \mid \boldsymbol{Z}\right\}=\boldsymbol{B}^{\top} \boldsymbol{Z}
$$

where

- $\hat{Z}\left(\mathbf{s}^{*}\right)$ are the interpolated Gaussian random field $\omega$ at location $\mathbf{s}^{*}$ which is the location of the block id,
- $\boldsymbol{B}$ is the cross-covariance matrix between the block id and survey stations,
- $\boldsymbol{Z}=\Sigma_{Z}^{-1} \times \bar{Z} . \Sigma_{Z}^{-1}$ is the inverse of the covariance matrix of the model with estimated spatial parameters (Tables 3.1 and 3.2) for $\omega$ based on the location of survey stations. $\bar{Z}$ is the estimated $\omega$ at each survey station.

Then the interpolation of the survey indices for target and non-target species can be done using

$$
\begin{gather*}
\hat{\lambda}_{T}\left(\mathbf{s}^{*}\right)=\exp \left(\beta_{T_{-} 0}+\hat{Z}_{T}\left(\mathbf{s}^{*}\right)\right)  \tag{3.1}\\
\hat{\lambda}_{N T}\left(\mathbf{s}^{*}\right)=\exp \left(\beta_{N T_{-} 0}+\hat{Z}_{N T}\left(\mathbf{s}^{*}\right)\right) \tag{3.2}
\end{gather*}
$$

Figure 3.6 is the interpolated random field and from this interpolation results we can obtain the interpolated survey indices for halibut (Figure 3.7). It can be observed from Figure 3.7 that there more halibut in the survey area $4 \mathrm{X} 5 \mathrm{YZ}, 4 \mathrm{~W}$ and 4 V .

| Year | Variance parameter $\sigma^{2}$ <br> for Target Species | Range Parameter $\phi$ <br> for Target Species | Smoothness Parameter $\nu$ <br> for Target Species |
| :---: | :---: | :---: | :---: |
| 2017 | 3.528151 | 0.031142 | 1 |
| 2018 | 3.162326 | 0.063860 | 1 |
| 2019 | 4.399422 | 0.062808 | 1 |

Table 3.1: Estimated parameters of Matérn covariance function for target species for 2017, 2018 and 2019. Smoothness parameter $\nu$ is fixed at one.

| Year | Variance parameter $\sigma^{2}$ <br> for Target Species | Range Parameter $\phi$ <br> for Target Species | Smoothness Parameter $\nu$ <br> for Target Species |
| :---: | :---: | :---: | :---: |
| 2017 | 1.538283 | 0.086193 | 1 |
| 2018 | 1.033545 | 0.06819973 | 1 |
| 2019 | 1.495082 | 0.05448649 | 1 |

Table 3.2: Estimated parameters of Matérn covariance function for non-target species for 2017, 2018 and 2019. Smoothness parameter $\nu$ is fixed at one.


Figure 3.6: Interpolated Gaussian random fields for halibut for 2017 (top), 2018 (middle) and 2019 (bottom). The black points are locations of the survey stations for each year.


Figure 3.7: The heatmaps of the interpolated relative abundance indices of halibut for 2017 (top), 2018 (middle) and 2019 (bottom).

### 3.3 Weighted Average of the Survey Indices

The aggregation of survey indices and the corresponding standard errors have been estimated using three different approaches so as to compare the changes in relative abundance indices of halibut over the past three years.

### 3.3.1 Explicit Method and corresponding Standard Error

The explicit method uses Equations 1.1 and 1.2 of the previous section to calculate the relative abundance indices for target and non-target species within each stratum. Then the weighted average for the entire survey area is obtained using the sum of the indices weighted by the corresponding stratum area and divided by the sum of the stratum area (Equation 3.1). In total, there are 15 strata with 3 depth strata and 5 area strata. However, stratum H13 is excluded from this analysis as the number of target species caught in 2017 for stratum H13 was 0 which would cause errors when calculating the survey indices. Bootstrapping is used to estimate the standard error of the estimated area weighted average of the survey indices [12]. Specifically, function bootstrap from R package bootstrap [25] is used to generate 15000 leave-one-out bootstrap replicates of Equations 1.1 and 1.2 applied to survey data. Then we calculate the standard error of observed value of Equations 1.1 and 1.2 applied to survey data.

$$
\begin{equation*}
\text { Area Weighted Average }=\frac{\sum_{i=1}^{n} A_{i} \lambda_{i}}{\sum_{i=1}^{n} A_{i}} \tag{3.3}
\end{equation*}
$$

where

- $A_{i}$ is the $i^{\text {th }}$ area of interest,
- $\lambda_{i}$ is the corresponding estimated relative abundance index for the $i_{t h}$ area,
- $n$ is the total number of area of interest.


### 3.3.2 Mean of the Interpolated Survey Indices and the Corresponding Standard Error

The aggregation of the survey indices via the interpolation method is over 1000 randomly selected locations from the block id. We use the built-in kriging function from

R package TMB to interpolate relative abundance indices over the 1000 randomly selected locations and then calculate the mean of the interpolated results. The standard error for this method is approximated using the delta method [6]:
let $\widehat{\boldsymbol{\theta}}_{n}$ be a sequence of $K \times 1$ random vectors such that

$$
\sqrt{n}\left(\widehat{\boldsymbol{\theta}}_{n}-\boldsymbol{\theta}\right) \xrightarrow{d} N(\mathbf{0}, \boldsymbol{\Sigma})
$$

where $N(\mathbf{0}, \boldsymbol{\Sigma})$ is a multivariate normal distribution with mean $\mathbf{0}$ and covariance matrix $\boldsymbol{\Sigma}, \boldsymbol{\theta}$ is a constant Kx 1 vector, and $\xrightarrow{d}$ indicates convergence in distribution. Let $\boldsymbol{g}: \mathbb{R}^{K} \rightarrow \mathbb{R}^{L}$. If all the $L$ entries of $\boldsymbol{g}$ have continuous partial derivatives with respect to $\boldsymbol{\theta}$, then

$$
\sqrt{n}\left(\boldsymbol{g}\left(\widehat{\boldsymbol{\theta}}_{n}\right)-\boldsymbol{g}(\boldsymbol{\theta})\right) \xrightarrow{d} N\left(\mathbf{0}, \nabla \boldsymbol{g}(\boldsymbol{\theta}) \boldsymbol{\Sigma} \nabla \boldsymbol{g}(\boldsymbol{\theta})^{\top}\right)
$$

where $\nabla \boldsymbol{g}(\boldsymbol{\theta})$ is the $L \times K$ matrix of partial derivatives of the entries of $\boldsymbol{g}$ with respect to the entries of $\boldsymbol{\theta}$.

### 3.3.3 Dirichlet Method and the Corresponding Standard Error

The Dirichlet tessellation method aggregates the survey indices at the level of the longline set. The Dirichlet tessellation of the survey stations is computed using the functions from the R package spatstat [4]. To be more specific, we use the function ppp from the $R$ package spatstat and the locations of survey stations to form the spatial point pattern and specify the enclosing polygon with the boundary points of survey area. The Dirichlet tile associated with a particular point i is the region of space that is closer to point i than to any other point in the spatial point pattern. Then we use function dirichlet from the R package spatstat to divide the survey area with Dirichlet tiles into disjoint regions, forming the tessellation (Figure 3.8). Finally we use the area of the Dirichlet tessellation and the corresponding estimated survey indices to calculate the weighted average (Equation 3.1). The delta method is used to estimate the standard error [6].


Figure 3.8: The Dirichlet tessellation and the survey stations for 2017 (top), 2018 (middle) and 2019 (bottom).

It can be observed from Figure 3.9 that the weighted relative abundance indices for halibut obtained using these three methods show a similar trend. The weighted survey indices increase from 2017 to 2018 and decrease from 2018 to 2019. The weighted survey indices obtained using the Dirichlet tessellation method are the largest when compared to the other two methods. Standard errors of Dirichlet tessellation method are the smallest among these three methods.


Figure 3.9: The weighted survey indices and corresponding standard errors for halibut obtained using three different weighting methods (all strata except H13) for 2017, 2018 and 2019.

## Chapter 4

## Discussion

Our results demonstrate (see Figures 3.1, 3.2 and 3.7) that catch rates for halibut are higher in shallow areas in southwest Nova Scotia (relative to shallow areas on Southern Grand Banks) and along shelf edges throughout the management unit. These conclusions are consistent with previous studies of Atlantic Halibut distribution. More recent presentations of commercial landings [11][15] also show similar patterns. Furthermore halibut catch rates in the DFO research vessel trawl surveys [8][15] showed juvenile hotspots and preferred habitat for halibut on shallow banks in southwest Nova Scotia (NAFO 4X and 4W, Strata 1) and along the shelf edge throughout the management unit.

We have shown that the differences between estimated survey indices for halibut obtained using fixed soak time and set-specific soak time are quite small when the explicit model is used. However, these differences are larger when the indices are estimated at the level of the longline set with our spatial model. It can be observed from Figures 4.1 through 4.3 that at a few survey stations, where the largest indices occur, the indices estimated using constant soak time are much smaller than the indices obtained using actual soak time. It also can be observed from these three plots that the estimates with actual soak time and mean soak time differ more when the estimated survey indices are large. Further investigation could be conducted to examine the causing of these large differences.

Although there are 1000 hooks sampled on a longline set, hook conditions are only observed for 300 of these hooks. This information is then used to estimate the effort of a longline set. For the remaining 700 hooks, the total number of halibut and non-target species caught are recorded. Earlier simulation analyses [24] suggest that 300 hooks are representative but the relative abundance indices for halibut estimated
using only this information are low. This suggests that the additional information from the remaining 700 hooks could be included in future analyses to obtain more precise estimates.


Figure 4.1: Relative abundance indices for halibut and corresponding standard errors estimated by models with set-specific soak time and mean soak time for 2017.


Figure 4.2: Relative abundance indices for halibut and corresponding standard errors estimated by models with set-specific soak time and mean soak time for 2018.


Figure 4.3: Relative abundance indices for halibut and corresponding standard errors estimated by models with set-specific soak time and mean soak time for 2019.

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## Appendix

## A. 1 Survey Indices Reproduction

The following codes are for reproducing the llsurv survey indices with the multinomial exponential model (MEM1) and R package TMB. The model is formulated in $\mathrm{C}++$ and the data are manipulated in R .

- $\mathrm{C}++$ Code for Model Formulation:

```
#include <TMB.hpp> // Links in the TMB libraries
template<class Type>
Type objective_function<Type>::operator() ()
{
    DATA_MATRIX(x); // The halibut data
    PARAMETER_VECTOR(theta); // Parameters
    DATA_VECTOR(s); // Soak time
    int n = x.col(0).size(); // Length of the column of matrix x
    int n_k = x.row(0).size(); // Length of the row of matrix x (4)
    // ldat - the relative abundance index for target species (halibut
        )
    vector<Type> ldat(n);
    for(int i=0; i<n; ++i){
        ldat(i)=exp(theta(0));
    }
    ADREPORT(ldat);
    // ldant - the relative abundance index for non-target species
    vector<Type> ldant(n);
    for(int i=0; i<n; ++i){
        ldant(i)=exp(theta(1));
    }
    ADREPORT(ldant);
    // pnt - the escaping rate for non-target species
    vector<Type> pnt(n);
```

```
27
28
29
30
31
- R Code for Data Manipulation:
```

\#\#For 2017

# Clear memory

rm(list=ls())

# The halibut data for 2017

data_2017 = read.csv("hdata2017.csv", header = T, stringsAsFactors =
F)
head(data_2017)

```
```

dim(data_2017)

# Corrected Strata ID

st_id = read.csv("HS_STATION_STRATA_2017_18_19.csv", header = F,
stringsAsFactors = F)
names(st_id) = c("YEAR", "STATION", "STRATA")
st_id_17 = st_id[which(st_id\$YEAR==2017), ]

# Merge the data

data2017 = merge(data_ 2017, st_id_17, by = "STATION")

# Drop the stations (523 and 525) with hook = 30

not30 = which(data2017$hooks_sampled == 30)
data2017[not30,]
data2017 = data2017[-c(not30), ]
which(data2017$hooks_sampled == 30)

## Data

# st is the soak time

# nb is the number of baited hooks at the end of the soak time

# nt is the number of individuals of the target species caught

# nnt is the number of individuals of the non-target species caught

# ne is the number of empty hooks at the end of the soak time

nb_17=data2017$empty_baited
nt_17=data2017$halibut
nnt_17=data2017$other_species
ne_17=data2017$empty_unbaited+data2017$broken_hook
st_17=data2017$SOAKMINP3P1

# The mean of the soak time (fix at the mean of the soak time)

st_17fixed=rep(mean(st_17), length(st_17))

## Calculate the initial values for TMB estimation

# Using the explicit equation to calculate the estimators (MEM1 with

    pt=0 from Etienne 2010)
    soak_17=mean(st_17)

# lambda for target species

```
```

ldathat_17=(sum(nt_17)/(sum(nt_17+nnt_17+ne_17+nb_17)-sum(nb_17)))*
(1/soak_17)*log(sum(nt_17+nnt_17+ne_17+nb_17)/sum(nb_17))
ldathat_17

# lambda non-target species

ldanthat_17=((sum(nnt_17)+sum(ne_17))/(sum(nt_17+nnt_17+ne_17+nb_17)
-sum(nb_17)))*(1/soak_17)*log(sum(nt_17+nnt_17+ne_17+nb_17)/sum(
nb_17))
ldanthat_17

# Escaping probabililty for target species

pthat_17=0
pthat_17

# Escaping probability for non-target species

pnthat_17=sum(ne_17)/(sum(ne_17)+sum(nnt_17))
pnthat_17

# Logit function

logitp=function(p){\operatorname{log}(p/(1-p))}

# Inverse logist function

logitpi=function(t){\operatorname{exp}(t)/(1+exp(t))}

## Use TMB to reproduce the results

library(TMB)

# Call TMB function value

compile("rep_code.cpp")

# Dynamically link the C++ code

dyn.load(dynlib("rep_code"))
data2017$empty= data2017$empty_unbaited+data2017\$broken_hook

# Put the required data into a matrix

x_17 = with(data2017, cbind(empty_baited, halibut, other_species,
empty))

# Soak time

st_17= as.vector(st_17)
st_17fixed = as.vector(st_17fixed)
6 9

```
```


# Data list

data_17= list(x=x_17,s=st_17)
data_17fixed = list(x=x_17,s=st_17fixed)

# Parameter list

param_17 = list()
param_17fixed = list()

# Initial values for lambda.t, lambda.nt and pnt

# Use the values calculated previously as the starting points

param_17$theta = c(log(ldathat_17), log(ldanthat_17), logitp(pnthat_
    17))
param_17fixed$theta = c(log(ldathat_17), log(ldanthat_17), logitp(
pnthat_17))

# Construct an R object (f) that represents our C++ function

# MakeADFun calls C++

f_17= MakeADFun(data_17, param_17, DLL="rep_code")
f_17fixed= MakeADFun(data_17fixed, param_17fixed, DLL="rep_code")

# Call TMB function value

fit_17= nlminb(f_17$par,f_17$fn,f_17$gr)
fit_17fixed= nlminb(f_17fixed$par,f_17fixed$fn,f_17fixed$gr)

# Calculate standard deviations of all model parameters

sdr_17= sdreport(f_17)

# Estimated results for model with actual soak time

summary(sdr_17)
sdr_17fixed = sdreport(f_17fixed)

# Estimated results for model with actual soak time

summary(sdr_17fixed)

```

\section*{A. 2 Binomial Model with Gaussian Random Field (glmmTMB)}

The following codes are used for estimating the probability of getting unbaited hooks over 1000 randomly selected survey stations from block id. R package glmmTMB and binomial model with Gaussian random field are used.
```


## For 2017

# Clear memory

rm(list=ls())
library(ggplot2)
library(dplyr)
library(marmap)

# The halibut data for 2017

data_2017 = read.csv("hdata2017.csv", header = T, stringsAsFactors =
F)
head(data_2017)
dim(data_2017)

# Corrected Strata ID

st_id = read.cSv("HS_STATION_STRATA_2017_18_19.csv", header = F,
stringsAsFactors = F)
names(st_id) = c("YEAR", "STATION", "STRATA")
st_id_17 = st_id[which(st_id\$YEAR==2017), ]

# Merge the data

data2017 = merge(data_2017, st_id_17, by = "STATION")
head(data2017)
dim(data2017)

# Drop the stations (523 and 525) with hook = 30

data2017 = data2017[-c(not30), ]
which(data2017\$hooks_sampled == 30)

# Read in the block id for the prediction

blockid = read.csv("blockIDkey.csv", header = T)
blockid[1:4, ]

# For binomial, we have baited and unbaited hooks (unbaited hooks +

    halibut + other species
    
# + broken hooks)

# Calulate the proportion unbaited hooks

```
```

data2017$biprop_unbaited = (data2017$hooks_sampled-data2017$empty_
        baited)/data2017$hooks_sampled
data2017$biprop_unbaited [1:10]
data17 = data2017 % >% dplyr::select(P1LONG.x, P1LAT.x, P1DEPTH,
    empty_baited, hooks_sampled, SOAKMINP3P1, MED_TEMP)
data17$biunbaited=data17$hooks_sampled - data17$empty_baited
d17 = data.frame(x = as.vector(data17$P1LAT.x),
    y = as.vector(data17$P1LONG.x))
library(glmmTMB)

# To fit the model, a numFactor and a dummy grouping variable must

        be added to the dataset:
    d17$pos <- numFactor(d17$x, d17$y)
d17$group <- factor(rep(1, nrow(d17)))
d17$soak = data2017$SOAKMINP3P1
d17$unbaited = data17$biunbaited
d17$baited = data17$empty_baited
d17$depth = data17$P1DEPTH

# With covariate temperature

mat_17 <- glmmTMB(cbind(unbaited,baited) ~ 1 + soak + depth + mat(
pos + 0 ( group),
family=binomial(link="logit"), data=d17)
summary(mat_17)

# Random sample }1000\mathrm{ points from the the block id without

        replacement for prediction
    samp = blockid[sample(nrow(blockid),size=1000,replace=FALSE),]
idpos = numFactor(samp$lat.DecDeg, samp$lon.DecDeg)

# Format the newdata which will be used for prediction

newdata_17 <- data.frame(pos=idpos, soak = mean(data2017$SOAKMINP3P1
    ), depth = samp$depth_m)
newdata_17\$group <- factor(rep(1, nrow(newdata_17)))
head(newdata_17)

```
```

6 1

# Predict the proportion of unbaited hooks

p_17=predict(mat_17, newdata_17, type="response", allow.new.levels=
TRUE)
head(p_17)

# Reshape the predicted proportion and corresponding lat and lon to

    a data.frame
    matn_17=data.frame(lati=samp$lat.DecDeg,
    long=samp$lon.DecDeg,
p=p_17)
write.csv(matn_17,file="matn_17.csv")

# Plot the predicted proportion of unbaited hooks

ggplot(matn_17, aes(x = long, y = lati, colour = p)) + geom_point()
+
scale_colour_gradient(low = "yellow", high = "red")

```

\section*{A. 3 Multinomial Exponential Model (MEM1) with Gaussian Random Field without Covariate (TMB)}

The following codes are used for estimating the relative abundance indices for target and non-target species at each survey station with spatial model without covariates and stratum H 13.
- \(\underline{\mathrm{C}++ \text { Code for Model Formulation: }}\)
```

\#include <TMB.hpp> // Links in the TMB libraries
template<class Type>
Type objective_function<Type>::operator() ()
{
DATA_MATRIX(H); // The halibut data
DATA_MATRIX(X); // Covariates matrix
PARAMETER_VECTOR(betat); // Parameters for lambda t (target
species)

```
```

PARAMETER_VECTOR(betant); // Parameters for lambda nt (non-target
species)
PARAMETER(theta); // Parameters for pnt (probability of caught non
-target fish escaping)
DATA_VECTOR(s); // Soak time
int n = H.col(0).size(); // Length of the column of matrix H
int n_k = H.row(0).size(); // Length of the row of matrix H (4)
PARAMETER_VECTOR(omegat); // The random field for lambda t
PARAMETER_VECTOR(omegant); // The random field for lambda nt
// ldat - the relative abundance indices for target species (
halibut)
vector<Type> ldat(n);
for(int i=0; i<n; ++i){
// Adding random field and covariates to lambda t
ldat(i)=exp((vector <Type>(X.row(i))*betat).sum()+omegat(i));
}
ADREPORT(ldat);
// ldant - the relative abundance indices for non-target species
vector<Type> ldant(n);
for(int i=0; i<n; ++i){
// Adding omega (random field) and covariates to lambda nt
ldant(i)=exp((vector <Type>(X.row(i))*betant).sum()+omegant(i));
}
ADREPORT(ldant);
// pnt - the probability of caught non-target species escaping
vector<Type> pnt(n);
for(int i=0; i<n; ++i){
pnt(i)=invlogit(theta);
}
ADREPORT(pnt);
DATA_MATRIX(D); // Distance matrix

```
```

PARAMETER(lognut);

```
PARAMETER(lognut);
PARAMETER(lognunt);
PARAMETER(lognunt);
PARAMETER(logPhit);
PARAMETER(logPhit);
PARAMETER(logPhint);
PARAMETER(logPhint);
PARAMETER(logSigmat);
PARAMETER(logSigmat);
PARAMETER(logSigmant);
PARAMETER(logSigmant);
// spatial parameters
// spatial parameters
Type nut=exp(lognut); // smoothness parameter
Type nut=exp(lognut); // smoothness parameter
Type nunt=exp(lognunt);
Type nunt=exp(lognunt);
Type phit=exp(logPhit); // range parameter
Type phit=exp(logPhit); // range parameter
Type phint=exp(logPhint);
Type phint=exp(logPhint);
Type sigt=exp(logSigmat); // variance
Type sigt=exp(logSigmat); // variance
Type signt=exp(logSigmant);
Type signt=exp(logSigmant);
// Covaraince matirx for random field for lambda t
// Covaraince matirx for random field for lambda t
matrix<Type> St(n,n);
matrix<Type> St(n,n);
St.setZero();
St.setZero();
for(int i=0; i<n; ++i){
for(int i=0; i<n; ++i){
    St(i,i) = sigt*sigt;
    St(i,i) = sigt*sigt;
}
}
for(int i=0; i<n; ++i){
for(int i=0; i<n; ++i){
    for(int j=i+1; j<n; ++j){
    for(int j=i+1; j<n; ++j){
            St(i,j) = sigt*sigt*matern(D(i,j), phit, nut); //Matern
            St(i,j) = sigt*sigt*matern(D(i,j), phit, nut); //Matern
    Covariance Function
    Covariance Function
            St(j,i) = St(i,j);
            St(j,i) = St(i,j);
    }
    }
}
}
// Covaraince matirx for random field for lambda nt
// Covaraince matirx for random field for lambda nt
matrix<Type> Snt(n,n);
matrix<Type> Snt(n,n);
Snt.setZero();
Snt.setZero();
for(int i=0; i<n; ++i){
for(int i=0; i<n; ++i){
    Snt(i,i) = signt*signt;
    Snt(i,i) = signt*signt;
}
}
for(int i=0; i<n; ++i){
```

for(int i=0; i<n; ++i){

```
```

        for(int j=i+1; j<n; ++j){
            Snt(i,j) = signt*signt*matern(D(i,j), phint, nunt); //Matern
        Covariance Function
            Snt(j,i) = Snt(i,j);
        }
    }
Type a = density::MVNORM(St)(omegat);
Type b = density::MVNORM(Snt)(omegant);
REPORT(St);
REPORT(Snt)
// Corresponding probabilities for N_b, N_t, N_nt and N_e
// N_b is the nmuber of baited hooks
// N_t is the number of target species caught
// N_nt is the number of non-target species caught
// N_e is the number of empty hooks
matrix<Type> p(n, n_k);
for(int i=0;i<n;i++){
p(i,0) = exp(-(ldat(i)+ldant(i))*s(i));
p(i,1)=(Type(1)-exp(-(ldat(i) +ldant(i))*s(i)))*(ldat(i)/(ldat(i)
+ldant(i)));
p(i, 2)=(Type(1)-exp(-(ldat(i) +ldant(i))*s(i)))*(ldant(i)/(ldat(i
)+ldant(i)))*(Type(1)-pnt(i));
p(i,3)=1-p(i,0)-p(i, 1)-p(i,2);
}
// The likelihood function for multinomial distribution
Type nll = 0;
for(int i=0; i < n; i++){
vector<Type> p_row=p.row(i);
vector<Type> H_row=H.row(i);
nll -= dmultinom(H_row, p_row,true);
}
nll=nll+a+b;

```
```

1 0 6
1 0 7
1 0 8
1 0 9

```
    REPORT(a);
```

    REPORT(a);
    REPORT(b);
    REPORT(b);
    REPORT(ldat);
    REPORT(ldat);
    REPORT(ldant);
    REPORT(ldant);
    REPORT(pnt);
    REPORT(pnt);
    REPORT(p);
    REPORT(p);
    REPORT(betat);
    REPORT(betat);
    REPORT(betant);
    REPORT(betant);
    REPORT(omegat);
    REPORT(omegat);
    REPORT(omegant);
    REPORT(omegant);
    return nll;
    ```
    return nll;
```

- R Code for Data Manipulation:
F)
head (data_2017)
dim(data_2017)
20

```
## For 2017
```


## For 2017

# Clear memory

rm(list=ls())
library(TMB)
library(fields)
library(dplyr)
library(ggplot2)
library(marmap)
library(RandomFields)

# Call TMB function value

compile("dmulticovariate_Oct.cpp")

# Dynamically link the C++ code

dyn.load(dynlib("dmulticovariate_Oct"))

# The halibut data for 2017

data_2017 = read.csv("hdata2017.csv", header = T, stringsAsFactors =

```
```

st_id = read.csv("HS_STATION_STRATA_2017_18_19.csv", header = F,
stringsAsFactors = F)
names(st_id) = c("YEAR", "STATION", "STRATA")
st_id_17 = st_id[which(st_id\$YEAR==2017), ]

# Merge the data

data2017 = merge(data_2017, st_id_17, by = "STATION")
head(data_ 2017)
dim(data2017)

# Drop the stations (523 and 525) with hook = 30

not30 = which(data2017$hooks_sampled == 30)
data2017[not30,]
data2017 = data2017[-c(not30), ]
which(data2017$hooks_sampled == 30)

# Drop the Stratum H13

h13_17 = which(data2017$STRATA == "H13")
data2017[h13_17,]
data2017 = data2017[-c(h13_17), ]
which(data2017$STRATA == "H13")
dim(data2017)

# Read in the block id for the prediction

blockid = read.csv("blockIDkey.csv", header = T)
blockid[1:4, ]
dim(blockid)

# Drop the Stratum H13

h13_id = which(blockid$s.id == 1.3)
blockid[h13_id,]
blockid = blockid[-c(h13_id), ]
which(blockid$s.id == 1.3)
dim(blockid)
library(sp)
library(rgdal)

```
```

library(INLA)
\#To switch for lat-long and project it on a flat surface since the
earth is a global
prj4s=CRS("+init=epsg:4326")
utm.prj4s=CRS("+init=epsg:32619")
loc_2017=cbind(data2017$P1LONG,data2017$P1LAT)
loc_17=cbind(data2017\$P1LONG, data2017 $P1LAT)
loc_17=SpatialPoints(loc_17, proj4string= prj4s)
loc_17=spTransform(loc_17,utm.prj4s)
data2017$P1LONG_proj=loc_17@coords[,1]
data2017\$P1LAT_proj=loc_17@coords[,2]

# Standardize loc for choosing a starting value for phi

spool_17= sqrt(((length(loc_2017[,1])-1)*var(loc_17@coords [, 1]) +(
length(loc_2017[, 2]) -1)*var(loc_17@coords [, 2]))/(length(loc_
2017[,1])+length(loc_2017[, 2])))
data2017$loc1=(loc_17@coords[,1]-median(loc_17@coords[,1]))/spool_17
data2017$loc2=(loc_17@coords[,2]_median(loc_17@coords[,2]))/spool_17
dismat_17=cbind(data2017$loc1,data2017$loc2)

# Distance matrix

Dist_17= as.matrix(dist(dismat_17))

# Put the required data into a matrix

data2017$empty=data2017$empty_unbaited+data2017\$ broken_hook
H_17=with(data2017, cbind(empty_baited, halibut, other_species,
empty))

# Soak time

s_17=data2017\$S0AKMINP3P1
s_17=as.vector (s_17)

# Logit function

logitp=function(p){\operatorname{log}(p/(1-p))}

# Inverse logist function

logitpi=function(t){\operatorname{exp}(t)/(1+exp(t))}
85

```
```


# The intercept (column of ones)

colone_17=rep(1, nrow(data2017))
X_170=as.matrix(colone_17)
data_170=list(H=H_17,s=s_17, X=X_170,D=Dist_17)
param_17tnt0 = list()

# Initial values for lambda.t, lambda.nt and pnt

# Use the estimated values as the starting points

param_17tnt0$betat = log(1.878483e-05)
param_17tnt0$betant = log(0.001941084)
param_17tnt0\$theta = logitp(0.8795768)

# Random field

param_17tnt0$omegat = rep(0, nrow(H_17))
param_17tnt0$omegant = rep (0, nrow(H_17))

# Smoothness parameter

param_17tnt0$lognut = 0
param_17tnt0$lognunt = 0

# Range parameter

param_17tnt0$logPhit = log(diff(range(dismat_17[,1]))*diff(range(
    dismat_17[,2])))
param_17tnt0$logPhint = log(diff(range(dismat_17[,1]))*diff(range(
dismat_17[,2])))

# Variance

param_17tnt0$logSigmat = 0
param_17tnt0$logSigmant = 0
newlist_170=list(lognut=factor(NA), lognunt=factor(NA))

# Construct an R object (f) that represents our C++ function

# use map and newlist debug

# Fix lognu at 0

f_17tnt0 = MakeADFun(data_170,param_17tnt0,random=c("omegat ","
omegant"), DLL="dmulticovariate_Oct",map=newlist_170, silent=TRUE)
fit_17tnt0 = nlminb(f_17tnt0$par,f_17tnt0$fn,f_17tnt0\$gr)

# Calculate standard deviations of all model parameters

sdr_17tnt0 = sdreport(f_17tnt0)

```
```

1 8
1 1 9
1 2 0
1 2 1
22
123
24
1 2 5
1 2 6
1 2 7
28
lamt_est_170 = data.frame(sum_sdr170[row.names(sum_sdr170) %in% "
ldat", ])
head(lamt_est_170)
data2017$lamt_se_170 = lamt_est_170$Std..Error
data2017$lamt_170 = lamt_est_170$Estimate

# For non-target species

head(sum_sdr170[row.names(sum_sdr170) %in% "ldant", ])
dim(sum_sdr170[row.names(sum_sdr170) %in% "ldant", ])
lamnt_est_170 = data.frame(sum_sdr170[row.names(sum_sdr170) %in% "
ldant", ])
head(lamnt_est_170)
data2017$lamnt_se_170 = lamnt_est_170$Std..Error
data2017$lamnt_170 = lamnt_est_170$Estimate

### Plot the Relative Abundance Indices

n_17 = c(1: nrow(data2017))
indices_17t = data2017$lamt_170
se_17t = data2017$lamt_se_170

# For target species

```
```

data_17t= data.frame(n=n_17, indices_t=indices_17t,se=se_17t)
p_17t = ggplot(data_17t, aes(x=n, y=indices_t)) + geom_point() +
geom_errorbar(aes(ymin=indices_t-se, ymax=indices_t+se), width
=0.5) + labs(x = "Survey Station", y = "Relative Abundance
Indices for Target Species")
p_17t

# For non-target species

indices_17nt = data2017$lamnt_170
se_17nt = data2017$lamnt_se_170
data_17nt = data.frame(n=n_17, indices_nt=indices_17nt,se2=se_17nt)
p_17nt = ggplot(data_17nt, aes(x=n, y=indices_nt)) + geom_point()+
geom_errorbar(aes(ymin=indices_nt-se2, ymax=indices_nt+se2),
width=0.5) + labs(x = "Survey Station", y = "Relative Abundance
Indices for Non-target Species")
p_17nt

```

\section*{A. 4 Interpolation of Estimated Relative Abundance Indices and Random Field}

The following codes are used for interpolating the random field and relative abundance indices for target species over the entire block id without stratum H13.
```


# Continuing from the code of MEM1 section

## For 2017

#### With R function

# For lambda t

# Range parameter for lambda t

logphit_170 = sum_sdr170 [row.names(sum_sdr170) %in% "logPhit", ][1]
logphit_170
phit_170 = exp(as.numeric(logphit_170))

# Variance for lambda t

logsigmat_170 = sum_sdr170[row.names(sum_sdr170) %in% "logSigmat",
] [1]
logsigmat_170
vart_170=(exp(as.numeric(logsigmat_170)))}\mp@subsup{)}{}{\wedge}

```
```


# Random field for lambda t

omt_170 = data.frame(sum_sdr170[row.names(sum_sdr170) %in% "omegat",
])
est_omt_170=omt_170$Estimate
#To switch for lat-long and project it on a flat surface since the
    earth is a global
loc_id=cbind(blockid$lon.DecDeg, blockid\$lat.DecDeg)
loc_idtrans=NULL
loc_idtrans=SpatialPoints(loc_id,proj4string = prj4s)
loc_idtrans=spTransform(loc_idtrans,utm.prj4s)

# Standardize block id

loc_idtransnorm_17=NULL
loc_idtransnorm_17$lon=(loc_idtrans@coords[,1]-median(loc_17@coords
    [,1]))/spool_17
loc_idtransnorm_17$lat=(loc_idtrans@coords[, 2]-median(loc_17@coords
[,2]))/spool_17
mloc_idtransnorm_17=as.matrix(cbind(loc_idtransnorm_17$lon,loc_
    idtransnorm_17$lat))

# Whittle-Matern Covariance Model

modelt_170 = RMwhittle(nu=1,scale=phit_170,var=vart_170)

# Interpolate omega (random field)

predt_170 = RFinterpolate(modelt_170,x=mloc_idtransnorm_17,
data=data.frame(x=dismat_17,data=est_omt_170),spConform=
FALSE)
lambdat_170=exp(as.numeric(sum_sdr170 [row.names(sum_sdr170) %in% "
betat", ][1])+predt_170)
blockid$intmega_17=predt_170
blockid$intldat_17=lambdat_170

# Plot the interpolated random field

col0=colorspace::diverge_hsv(256)
col2=c(col0[c(seq(1,127,by=3),128,129:256)])

```
```


# Random field

quilt.plot(loc_id, matrix(predt_170), xlab="long",ylab="lati",main=
expression(paste("Heat map of random field for ", lambda,"_t for
2017')), nx=220, ny=220, col=col2, zlim=c(-1.4,4.2))
world(add=TRUE, col="black", lwd=1)
points(loc_2017, type="p",pch=20, cex=0.6)

# Plot the interpolated relative abundance indices

col3=c(col0[c(128,129:256)])
quilt.plot(loc_id,matrix(lambdat_170),main=expression(paste("
Estimated ", lambda, "_t for 2017")),xlab="long",ylab="lati",col=
col3, zlim=c(0,max(lambdat_170)))
world(add=TRUE, col="black", lwd=1)
points(loc_2017, type="p",pch=20, cex=0.6)

#### With Theoretical Calculation

# Covariance matrix for Whittle-Matern model with 2017 survey

    location
    covy=RFcovmatrix(modelt_170,x=dismat_17)

# Covariance matrix for Whittle-Matern model with block id location

covy0=RFcovmatrix(modelt_170,x=mloc_idtransnorm_17)

# Distance between block id location and 2017 survey location

disy0y=rdist(mloc_idtransnorm_17,dismat_17)

# Computes the empirical cross-covariance function for given spatial

    data.
    
# Covariance for Whittle-Matern model with block id location

# and Whittle-Matern model with 2017 survey location

covy0y=RFcov(modelt_170,dist=as.vector(disy0y), dim=2)
covy0y=matrix(covy0y, ncol=nrow(covy))

# Interpolate omega (random field)

y0=covy0y%*%solve(covy, est_omt_170)

# Check R function result and theoretical calculation results

plot(y0,predt_170)
sum((y0-predt_170) ^2)

```

\section*{A. 5 Aggregation with Dirichlet Method}

The following codes are used for calculating weighted average of relative abundance indices for halibut using Dirichlet method without stratum H13.
```


# Continuing from the code of MEM1 section

## For 2017

## Dirichlet

library(gissr)
library(deldir)
library(spatstat)
library(alphahull)

# Read in block id

bid = read.csv("blockIDkey.csv", header = T)
loc_id = cbind(bid$lon.DecDeg, bid$lat.DecDeg)

# Plot block id

plot(loc_id, xlab="Long", ylab="Lat", main="blockIDkey", pch=20)

# With Alpha Hull

# Find the boundary points using ashape

# Alpha controls the detail of the boundary

bound_a = ashape(loc_id, alpha = 0.084)
bound_a_index = bound_a\$alpha.extremes
plot(loc_id, xlab="Long", ylab="Lat", main="blockIDkey and boundary"
, pch=20)
points(loc_id[bound_a_index, ], col=2, pch=20)

### alpha = 0.084

# Boundary points and survey stations

bound_a_pos = loc_id[bound_a_index, ]
plot(bound_a_pos, pch=20)

# Extracted block id boundary and imported boundary

plot(bound_a_pos, xlab="Long", ylab="Lat", main=" Survey Area
Boundary Coordinate", pch=20)
points(surveyPolyLL$X, surveyPolyLL$Y, col="red", pch=20)

```
```

30

# To find the corresponding index

bound_a_pos = data.frame(long=bound_a_pos[,1], lat=bound_a_pos [, 2])
bound_a_pos$num = 1:length(bound_a_pos$long)

# Fix the boundary

dim(bound_a_pos)
bound_a_pos = bound_a_pos[-c (12, 15:21,57, 58,579,581,588,594,
595,598,599,601,604,605,668,669,828,834,840,844,845,841,835,
773,751,755,320,311:313,326,340,357,361,364,365,360,356,353,
336,325,452,472,495,507,678,685,686,634,630,635,639,638,640,
323,316,306),]
dim(bound_a_pos)
plot(bound_a_pos$long, bound_a_pos$lat, pch=20)

# To switch for lat-long for block id

# and project it on a flat surface since the earth is a global

tf_bound_a_pos = SpatialPoints(bound_a_pos, proj4string = prj4s)
tf_bound_a_pos= spTransform(tf_bound_a_pos,utm.prj4s)
tf_bound_a_pos = data.frame(long = tf_bound_a_pos@coords[,1], lat =
tf_bound_a_pos@coords[,2])
plot(tf_bound_a_pos)

# List the boundary points in anit-clockwise order

tf_bound_a_pos2 = sort_points(tf_bound_a_pos, "lat", "long",
clockwise = FALSE)
tf_bound_a_pos2 = data.frame(long = tf_bound_a_pos2$long, lat = tf_
        bound_a_pos2$lat, num=1:length(bound_a_pos\$long))

# Creates an object of class "ppp" representing a point pattern

        dataset in the two-dimensional plane
    pp_17=ppp(data2017$P1LONG_proj, data2017$P1LAT_proj, window=owin(
poly=list(x=tf_bound_a_pos2$long, y=tf_bound_a_pos2$lat)))
dpp_17 = dirichlet(pp_17)

```
```


## Dirichlet plot

tf_bound_a_pos22=tf_bound_a_pos2[926, ]
tf_bound_a_pos22[1:64, ] =
tf_bound_a_pos2[c(918, 921,925,929,932,936,935,934,933,931,
930, 927, 926, 924, 923, 922, 920, 919, 928, 937:953, 955, 954, 956,
958,960, 957, 959,40,48,55,61,59,66,67,76,84, 87, 94,95,91,
97,99,101,102,105,106,110,124), ]
tf_bound_a_pos22[65:160, ] =
tf_bound_a_pos2[c(127,117,120,123,113,116,118,122,125,115,
129,134,138,142,148,144,140,136,131,111,119,128,141, 146,
150,152,156,158,162,163,166,167,170,176,179,181,185,186,
189,190,194,195,191,196,199,201,202, 205,215, 214, 213,212,
207, 211, 210, 209, 206, 204, 198,192,187,182,177,174,171, 164,
161,154, 151, 145,135,133,130,112,108,107,104,96,92, 90, 89,
83,80,77,75,70,68,64,62,60,57,54,52,46,43,37),]
tf_bound_a_pos22[161:250, ] =
tf_bound_a_pos2[c(36, 27, 29, 28, 23,17,14,12,10,13,15,19, 22,
20,18,6,5,2,963,961,962,1,3,4,7:9,11, 16, 21, 24, 25, 26, 33, 35,
32, 30, 31, 34, 38, 39, 42, 45,49, 41,44, 47,50,51,53,56,58,63,65,
69,73,81,86,78,71,72,74,79,82,85,88,93,98,100,103,109,114,
121,126,132,137,139,143,147,149,153,155,157,159,160,165,
168,169,172,173), ]
tf_bound_a_pos22[251:371, ] =
tf_bound_a_pos2[c(175,178,180,183,184,188,193,197,200, 203,
208, 216:235, 250, 260, 265, 272, 280, 283, 294, 295, 292, 290, 286,
285,288, 287, 281, 278, 277, 276, 274, 271, 270, 269, 266, 261, 254,
247, 245, 240, 236, 241, 237, 242, 238, 243, 248, 244, 239, 246, 251,
255,252, 256, 253, 249, 257:259, 262:264, 267, 268, 273, 275,279,
282, 284, 289, 291, 293,296:306,309,307, 308, 310:321, 323,328,
332,331), ]
tf_bound_a_pos22[372:444, ] =
tf_bound_a_pos2[c
(327, 325,322, 324, 330, 326, 329, 333:349,352, 351, 350, 353:398), ]
tf_bound_a_pos22[445:545, ] =
tf_bound_a_pos2[c(401, 400, 399,402:405,426,447,460,469,486,

```
```

    491,512,517,522,530,523,531,524,519,516,526,535,540,549,
    557,559,555,563,571,576,577,574,560,562,570,572,567,564,
    566,569,558,551,548,543,545,542,544,541,538,537,532,527,
    533,528,525,521,518,508,514,510,506,511,507,504,499,505,
    501,496,502,497,503,498,493,487,483,480,484,482,479,477,
    476,474,473,470,466,461,458,457,463,459,455,453,445,444,
    449,446,451,450,441), ]
    tf_bound_a_pos22[546:615, ] =
tf_bound_a_pos2[c(439,437,443,440,438,436,427,409,406,407,
413,431,435,434,424,420,428,425,422,418,415,411,408,417,
414,410,421,416,412,423,419, 432, 430,429,433,442,448,452,
454,456,462,464,465,467,468,471,472,475,478,481,485,488,
489,490,492,494,495,500,509,513,515,520,539,534,536,539,
546,547,550,552), ]
tf_bound_a_pos22[616:688, ] =
tf_bound_a_pos2[c(553,554,556,561,565,568,573,575,578:612,
620,619,618,617,616,615,614,613,627,626,625,624,623,622,
621,633,632,631,630,629,628,639,638,637,636,635,634,645,
644,643), ]
tf_bound_a_pos22[689:785, ] =
tf_bound_a_pos2[c (642,641,640,650,649,648,647,646,656,655,
654,653,652,651,657:739), ]
tf_bound_a_pos22[786:874, ] =
tf_bound_a_pos2[c(750,756,764,769,780,777,776,773,770,779,
766,765,761,754,752,749,746,744,751,757,760,768,778,775,
772,767,763,759,758,747,745,742,741,740,743,753,748,755,
762,774,771,781,783,782,784:786,790,803,815:818,811,810,
806,808,798,800,802,792,794:796,799,804, 805,807,797,791,
787:789,793,801, 809, 812, 819, 822, 824, 825, 823, 814, 813,821,
820,826,828,827), ]
tf_bound_a_pos22[875:963, ] = tf_bound_a_pos2[c
(829:913,915,914,916:917), ]

```
tf_bound_a_pos22 = data.frame (long = tf_bound_a_pos22\$long, lat \(=t f\)
        _ bound_a_pos22\$lat, num=1: length(tf_bound_a_pos22\$long))
```

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1 2 8

# Creates an object of class "ppp" representing a point pattern

        dataset in the two-dimensional plane
    pp_17=ppp(data2017$P1LONG_proj, data2017$P1LAT_proj, window=owin(
poly=list(x=tf_bound_a_pos22$long, y=tf_bound_a_pos22$lat)))
dpp_17 = dirichlet(pp_17)
plot(dpp_17,border="black",main="")
plot(pp_17, add=TRUE, chars=20, col="red", main="Dirichlet Tiles Plot
for 2017")

## Area

wts_17 = sapply(tiles(dpp_17),area.owin)
cat("Sum of weights:\n")
print(sum(wts_17)/1000000)

# Target

dweight_ave_17t = sum((data2017\$lamt_170)*wts_17)/sum(wts_17)
dweight_ave_17t

# Standard error calculation

wts_17 = as.matrix(wts_17)

# wi/sum(wi)

wts_172 = wts_17/(sum(wts_17))

# covariance matrix for estimated lambda t

cov_17= sdr_17tnt0\$cov
cov_17t = cov_17[1:nrow(data2017), 1:nrow(data2017)]

# plot((data2017\$lamt_se_170) ^2, diag(cov_17t))

sum(((data2017\$lamt_se_170) ^2-diag(cov_17t)))

# Standard error for Dirichlet method

se2_dir_17t = t(wts_172)%*%cov_17t%*%wts_172
sqrt(se2_dir_17t)

```
```

